

Threatened pond endemicity on an oceanic island: the presence of an exotic fish

Margarita Florencio^{1,2,3,*} and Lucas Lamelas-López^{1,2}

¹ CE3C-Centre for Ecology, Evolution and Environmental Changes. Azorean Biodiversity Group. Universidade dos Açores, Departamento de Ciências Agrárias, Angra do Heroísmo, Açores, Portugal.

² CITA-A. Universidade dos Açores, Departamento de Ciências Agrárias, Angra do Heroísmo, Açores, Portugal.

³ Present address: Departamento de Ecologia, Universidade Federal de Goiás, 74001-970, Goiânia, Goiás, Brazil.

* Corresponding author: margarita@ebd.csic.es; mflorenciodiaz@gmail.com

Received: 06/08/2014

Accepted: 21/12/2015

ABSTRACT

Threatened pond endemicity on an oceanic island: the presence of an exotic fish

The EU Water Framework Directive (WFD) prioritises the importance of freshwater conservation areas in the Macaronesian Islands (Azores, Madeira, Canarias and Cape Verde). However, few studies have investigated these freshwater areas. Therefore, research must focus on establishing reference sites for aquatic conservation. We analyse the potential effects of water characteristics and the presence of an exotic predator fish, *Gambusia holbrooki* (Girard, 1859), on the abundance of an endemic aquatic beetle, *Hydroporus guernei* (Régimbart, 1891), on an oceanic island. During ten successive months, we sampled the abundance of *H. guernei* using a protocol that targeted macroinvertebrates in three temporary and two permanent ponds on Terceira Island, Azores archipelago. We also recorded the abundance of *G. holbrooki*, which was observed for the first time on the island, as well as the environmental characteristics of these ponds. *H. guernei* contributes a high degree of endemicity to the temporary ponds, while the exotic species only occurred in one permanent pond. The results of multiple regressions suggest that the abundance of *H. guernei* can be mainly explained by the conductivity (19.1%) and the abundance of *G. holbrooki* (14.5%). These results suggest that *H. guernei* is a species with a narrow tolerance to environmental changes. In addition, its abundance may be affected by the presence of *G. holbrooki*. We suggest that urgent monitoring actions be taken to control the abundance of this exotic fish in these ponds, thereby guaranteeing the conservation of endemic *H. guernei* populations.

Key words: Acidic waters, Azores, endemic species, *Gambusia holbrooki*, *Hydroporus guernei*, aquatic macroinvertebrates, temporary ponds.

RESUMEN

Amenazada la endemicidad de las lagunas en una isla oceánica: la presencia de un pez exótico

La Directiva Marco del Agua (WFD) de la UE destaca la importancia para la conservación de los humedales en las islas macaronésicas (Azores, Madeira, Canarias y Cabo Verde). Sin embargo, sólo unos pocos estudios han considerado estos humedales, siendo esencial llevar a cabo estudios enfocados en establecer puntos de referencia para la conservación acuática. En este estudio analizamos el posible efecto de las características del agua y de la presencia de un pez exótico depredador, *Gambusia holbrooki* (Girard, 1859), sobre la abundancia de un coleóptero acuático endémico, *Hydroporus guernei* (Régimbart, 1891), en una isla oceánica. Durante diez meses sucesivos, muestreamos la abundancia de *H. guernei* usando un protocolo dirigido al muestreo de macroinvertebrados, en tres lagunas temporales y dos permanentes, en la isla Terceira, archipiélago de las Azores. También muestreamos la abundancia de *G. holbrooki*, observada por primera vez en la isla de estudio, y las características ambientales de estas lagunas. Los individuos de *H. guernei* atribuyeron un alto grado de endemicidad a las lagunas temporales, mientras que el pez exótico solo apareció en una laguna permanente. Después de realizar regresiones múltiples, detectamos que la abundancia de *H. guernei* fue principalmente explicada por la conductividad (19.1%) y por la abundancia de *G. holbrooki* (14.5%). Estos resultados sugieren que *H. guernei* es una especie con una estrecha tolerancia ambiental, y que su abundancia se puede ver afectada por la ocurrencia de *G. holbrooki*. Sugérimos que

se deberían llevar a cabo acciones urgentes de monitoreo en estas lagunas para controlar la abundancia de este pez exótico, y garantizar así la conservación de las poblaciones endémicas de *H. guernei*.

Palabras clave: Aguas ácidas, Azores, especies endémicas, *Gambusia holbrooki*, *Hydroporus guernei*, macroinvertebrados acuáticos, lagunas temporales.

INTRODUCTION

The EU Water Framework Directive (WFD) prioritises freshwater conservation areas in the Macaronesian region, which comprises Azores, Madeira, Canarias and Cape Verde Islands. The high level of endemism associated with the Macaronesian macroinvertebrate fauna creates considerable conservation value in these aquatic habitats. However, the lack of information at reference sites hinders the establishment of proper conservation guidelines (Hughes & Malmqvist, 2005). Particularly, temporary ponds, which undergo re-

current inundation and unpredictable desiccation, are considered priority habitats for conservation by the European Union (code 3170 of the Habitat Directive). Temporary ponds harbour singular species. These species must cope with pond desiccation by adjusting their life cycles based on the variable duration of pond inundation (Williams, 2006). In contrast to permanent ponds, temporary ponds are characterised by the absence of top-predator fishes, which cannot cope with pond desiccation (Wellborn *et al.*, 1996). The introduction of exotic species has contributed to declining global biodiversity (Vitousek *et al.*, 1996,

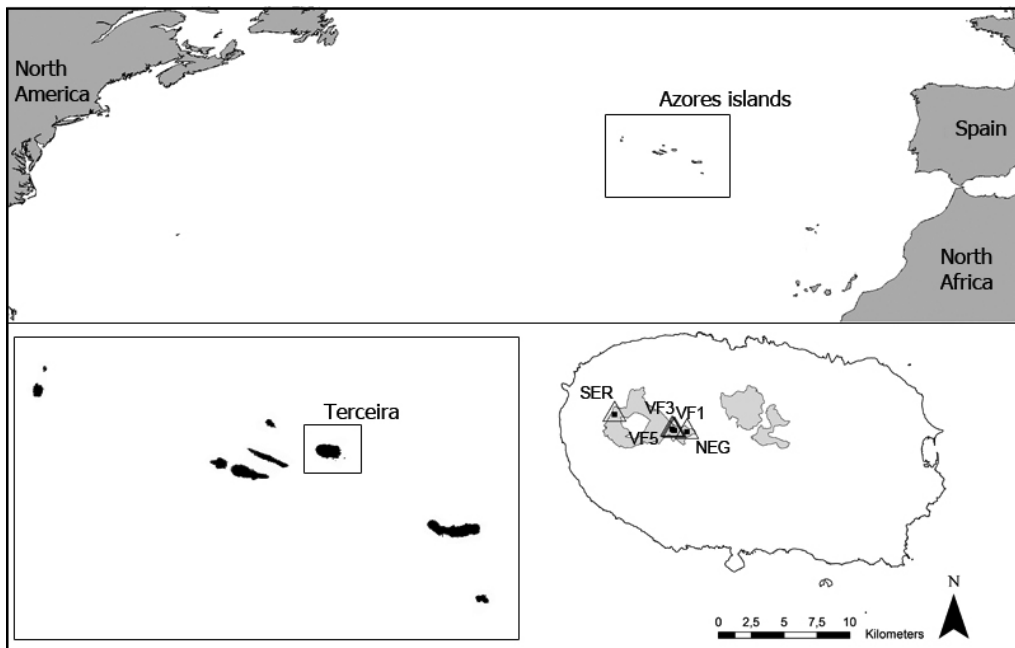


Figure 1. Map of the Azores archipelago showing the five study ponds on Terceira Island (VF1, VF3 and VF5 are temporary ponds, and NEG and SER are permanent ponds). The Natural Forest Reserve (NFR) in “Serra de Santa Bárbara and Mistérios Negros” (grey area) and the location of the Terceira Island in the Azores archipelago are also indicated. *Localización del archipiélago de las Azores y de las cinco lagunas de estudio en la isla Terceira (VF1, VF3 y VF5 son temporales, y NEG y SER son permanentes); también se indica la Reserva Natural Forestal (NFR) de “Serra de Santa Bárbara e Mistérios Negros” (área gris) y la localización de la isla Terceira en el archipiélago de las Azores.*

1997; Cassey *et al.*, 2005). This biodiversity loss has extended to aquatic ecosystems, resulting in changes to biodiversity patterns (Hermoso *et al.*, 2012) and habitat quality degradation (Arribas *et al.*, 2014). Specifically, the introduction of exotic species at high trophic levels may promote accentuated changes on oceanic islands where native biotas have often evolved without predators (Whittaker & Fernández-Palacios, 2007).

In our study, we analysed the abundance of the aquatic beetle *Hydroporus guernei* (Régimbart, 1891), which has been described as an endemic species of the Azores archipelago by Alluaud (1891), in three temporary and two permanent ponds on Terceira Island. We analysed the effects of an exotic predator fish, *Gambusia holbrooki* (Girard, 1859), observed for the first time on the study island, versus other environmental pond characteristics. We hypothesised that the predator fish would strongly affect the abundance of the endemic beetle in permanent ponds compared to temporary ponds, which would be less susceptible to the introduction of the exotic fish.

METHODS

Study site

The Azores archipelago is located between 36° 45' – 39° 43' N and 24° 45' – 31° 17' W. It comprises nine islands separated in three groups: the western group (Flores and Corvo), the central group (Faial, Pico, S. Jorge, Terceira and Graciosa) and the eastern group (S. Miguel and S. Maria), in addition to small islets. The temperate oceanic climate is characterised by stable temperatures, substantial precipitation and high relative atmospheric humidity, which can reach more than 95% in high-altitude native forests. This study was conducted between November 2013 and August 2014 in three temporary ponds, VF1, VF3 and VF5, and two permanent ponds, Negro (NEG) and Serreta (SER) (Fig. 1; the UTM geographical coordinates X, Y in zone 26S are 475712, 4287838 for VF1; 475605, 4287864 for VF3; 475505, 4287935 for VF5; 476639, 4287739 for NEG; and 471127, 4289052 for

SER). Among the temporary ponds, VF1 was dry in June, VF5 was dry in July and VF5 contained water throughout the sampling period; although, it was dry in October before the study began. The ponds were all located in the Natural Forest Reserve (NFR) created under Portuguese law n° 27/88/A within the Natural Park of Terceira in Serra de Santa Bárbara and Mistérios Negros. This NFR encompasses the majority of the native forest area on Terceira Island (Fig. 1), currently accounting for 6% of the island area.

Pond characteristics and sampling procedures

Study sites were sampled monthly. Electrical conductivity at 20 °C (Multi-range Conductivity meter HANNA HI 98127), pH (pH meter HANNA HI 98311), dissolved O₂ concentration and temperature (°C) (Oxi 315i WTW) were measured *in situ* in the water column. We obtained two replicates of each measurement per pond and per month. An average value was used given the homogeneity of the measures. We collected 330 ml of surface water to measure the total P concentration in the water column (TP) using acid digestion in the laboratory (Golterman, 2004). We sampled individuals of *Hydroporus guernei* (adults and larvae) using a sampling procedure that targets macroinvertebrates. The method utilized a 40 × 25 cm dip net with a 1 mm mesh size. Approximately 1.5 m stretches of water were netted in each sampling unit. We sampled all available microhabitats in each pond based on differences in aquatic plant cover and depth (Heyer *et al.*, 1994), extending from the littoral zone to the deepest point of the pond when accessible. The dip netting efficiency was generally better in small ponds (Heyer *et al.*, 1994). Therefore, we assigned a specific number of sampling units to each pond, with larger numbers given to larger ponds (Florencio *et al.*, 2009). The study consisted of 557 total sampling units, ranging from 5 to 19 sampling units per pond. Species sorting was performed in the field. Most adult individuals of *H. guernei* were released back into the pond. Only unidentified larvae were preserved in 70% ethanol for subsequent identification in the laboratory. Using the described

sampling procedure that targets macroinvertebrates, we also captured individuals of *Gambusia holbrooki*, which were identified and quantified.

Data analyses

We pooled adults and larvae together as individuals of *H. guernei*. We then calculated the average number of individuals captured per pond and per month (hereafter called “abundance of *H. guernei*” for simplicity). The average number of individuals of *G. holbrooki* per pond and per month was also calculated (hereafter called “abundance of *G. holbrooki*” for simplicity). Because we could not record the dissolved O₂ concentration in February and March, we replaced missing values with the pond-based mean of that variable (e.g., Leps & Smilauer, 2003). The abundance of *H. guernei* was $\log_e(x + 1)$ transformed to obtain normality. This parameter was then used as a response variable in the statistical analyses. We then used the abundance of *G. holbrooki* and the pond characteristics as predictors using general linear models for multiple regressions (Legendre & Legendre, 1998; Statistica v. 8). To avoid collinearity, we excluded pH from the statistical analysis due to the correlation with the abundance of *G. holbrooki* (Spearman correlation = 0.76). Type III sums of squares were used and standardised β -coefficients were estimated to measure both the relative contribution of each variable (electrical conductivity, dissolved O₂ concentration, total P concentration in the water column and the abundance of *G. holbrooki*) and their relationships to the abundance of *H. guernei*.

RESULTS AND DISCUSSION

Pond and biological characteristics

The ponds were characterised as well-oxygenated waters (Table 1). Low water conductivities (<109 $\mu\text{S}/\text{cm}$) indicate that the ponds were primarily flooded with rainwater. A reduced TP was detected in the water column, likely due to water dilution as a consequence of persistent rainfall.

TP only peaked in a permanent pond (NEG) during June. This result may have been due to the combination of a low water level and windy days (pers. observation), which favour the suspension of particles from the sediment. Rainfall may have also contributed to the acidic waters detected in all ponds across the sampling months, resulting in a pH of <4.2 in temporary ponds (Table 1). However, other factors may have contributed to such low pH values in the study ponds, e.g., the presence of basalt deposits of volcanic origin (Timperley & Vigor-Brown, 1986; Delmelle & Bernard, 1994; Bernard *et al.*, 2004), humic substances washed from upland soils into the study ponds (Florencio *et al.*, 2013) and secretions by the organic acidic molecules in *Sphagnum* (Clymo, 1964; Mendes & Dias, 2001).

Adults and larvae of *H. guernei* were observed simultaneously during the sampling period, peaking in abundances during three different months (Table 1). The abundance variations are likely associated with annual rainfall patterns. In November, the temporary ponds had recently been inundated; however, precipitation was still low compared to December. In February, after two successive months with little rainfall, the water level reduction may have concentrated high abundances of *H. guernei*. In May, ponds also exhibited reduced water levels due to minimal rainfall, followed by desiccation in two temporary ponds in the successive months.

Pond endemism

In total, we detected 266 individuals (176 adults and 90 larvae) of *H. guernei* in the ten months of sampling (Table 1), highlighting the high endemism of the Azorean ponds (e.g., Raposeiro *et al.*, 2012). In temporary ponds, we recorded 0.9 (± 1.9 SD) individuals of *H. guernei* on average per pond and per month, while only 0.1 (± 0.3 SD) individuals on average were detected in the permanent ponds. This result suggests that *H. guernei* contributes a high degree of endemism to the temporary ponds. In this study, we have reported the introduction of the mosquitofish *Gambusia holbrooki* (Girard 1859) for the first time on Terceira Island. This invasive species

Table 1. Average number of *H. guernei* individuals per pond based on adults and larvae. The average number of *Gambusia holbrooki* individuals per pond and the environmental characteristics of the temporary and the permanent ponds are also indicated (the blanks indicate the dissolved oxygen concentration values that were not measured; NEG and SER are the permanent ponds; VF1, VF3 and VF5 are the temporary ponds; VF1 was dry in June; VF1 and VF5 were dry in July and August). *Número medio de individuos de H. guernei por laguna de adultos y de larvas. También se indica el número medio de individuos de Gambusia holbrooki por laguna y las características ambientales de las lagunas temporales y permanentes (los espacios en blanco indican aquellos valores de la concentración de oxígeno disuelto que no se midieron; NEG y SER son las lagunas permanentes y VF1, VF3 y VF5 son las lagunas temporales; VF1 aparece seca en Junio, y VF1 y VF5 también en Julio y Agosto).*

Month	Pond	<i>H. guernei</i> Adult	<i>H. guernei</i> Larvae	Averaged abundance of <i>Gambusia</i>	TP (µg/L) ^a	O ₂ (mg/L) ^b	pH	Conductivity (µS/cm)
November								
	NEG	0	0	2.57	66.90	4.23	6.55	44.50
	SER	0.08	0.08	0	10.46	4.58	5.68	45.38
	VF1	1.13	1.38	0	36.00	5.08	5.08	55.25
	VF3	0.22	0.11	0	10.22	4.00	4.85	49.25
	VF5	0.60	0.70	0	19.70	5.08	4.23	40.75
December								
	NEG	0	0	1.00	23.11	5.60	6.00	18.00
	SER	0	0	0	5.60	5.08	5.10	29.50
	VF1	0.57	0	0	21.16	4.83	5.00	31.00
	VF3	0	0.22	0	5.11	4.98	4.70	27.00
	VF5	0.20	0	0	12.41	5.53	4.45	26.50
January								
	NEG	0	0	4.61	15.57	6.23	5.90	44.25
	SER	0.07	0	0	20.92	5.83	5.23	40.50
	VF1	0.13	1.00	0	19.70	6.25	5.03	56.00
	VF3	0	0	0	10.95	6.20	4.53	51.25
	VF5	0.50	0	0	9.24	6.23	4.15	55.75
February								
	NEG	0.21	0	3.42	18.00		6.05	63.00
	SER	0	0	0	14.60		5.13	59.50
	VF1	2.00	1.55	0	17.27		4.90	108.00
	VF3	0.63	0.38	0	4.87		4.45	74.75
	VF5	0.60	0.20	0	9.73		4.10	83.25
March								
	NEG	0	0	2.19	23.84		6.25	36.00
	SER	0.06	0.00	0	19.22		5.15	54.50
	VF1	1.21	0.21	0	27.49		5.08	37.25
	VF3	0	0	0	11.68		4.70	48.00
	VF5	0.58	0.08	0	27.25		4.40	39.75
April								
	NEG	0	0	2.82	27.49	7.60	6.30	40.50
	SER	0.19	0.06	0	13.14	7.05	5.15	56.00
	VF1	0.80	0	0	18.73	6.68	5.20	45.75
	VF3	0.08	0.15	0	13.38	7.05	4.83	43.75
	VF5	0.45	0.09	0	25.54	7.78	4.55	37.00
May								
	NEG	0.06	0	4.83	29.37	7.00	6.95	37.75
	SER	0.35	0	0	29.13	7.58	5.50	47.75
	VF1	2.50	2.00	0	36.41	8.25	5.10	36.75
	VF3	0.17	0	0	25.49	6.85	4.90	41.00
	VF5	1.18	0.09	0	26.94	7.93	4.33	68.50
June								
	NEG	0	0	3.08	85.14	5.90	6.90	43.00
	SER	0	0.07	0	1.46	6.80	5.73	45.75
	VF1							
	VF3	0	0	0	12.89	6.20	5.20	37.00
	VF5	1.00	0	0	23.60	9.50	4.45	79.00
July								
	NEG	0.08	0	5.67	14.84	7.03	6.50	46.00
	SER	0	0	0	5.60	6.95	5.60	43.00
	VF1							
	VF3	0	0	0	13.87	6.35	5.45	38.00
	VF5							
August								
	NEG	0	0	5.20	11.19	7.15	6.30	30.50
	SER	0	0	0	5.35	6.90	5.70	36.25
	VF1							
	VF3	0	0	0	10.46	5.20	5.55	26.50
	VF5							

^a is the total P concentration in the water column; ^b is the dissolved oxygen concentration in the water column.

Table 2. Variability (Var %) in the abundance of *H. guernei* due to electrical conductivity (EC), dissolved O₂ concentration (O₂), the total P concentration in the water column (TP) and the abundance of *G. holbrooki* (*Gambusia*). Beta-coefficients (β), which indicate the sign of the effect (positive or negative), the standard errors of the Beta-coefficients (SE (β)), the statistical values (*F*), the degrees of freedom (Df.) and the *P*-values are also indicated. *Variabilidad pura (Var %) de la abundancia de H. guernei explicada por la conductividad eléctrica (EC), la concentración de O₂ disuelto (O₂), la concentración del P total en la columna de agua (TP) y la abundancia de G. holbrooki (Gambusia). Se indican los Beta-coeficientes (β), indicando el signo de cada efecto (positivo o negativo), el error estándar de los Beta-coeficientes (SE (β)), el valor del estadístico (F), los grados de libertad (Df.) y los P-valores.*

	Df.	<i>F</i>	Var %	<i>P</i> -value	β	SE (β)
EC	1	1.18	19.1	0.004	0.41	0.13
O ₂	1	0.10	1.9	0.381	0.12	0.13
TP	1	0.44	8.1	0.067	0.25	0.13
<i>Gambusia</i>	1	0.84	14.5	0.013	-0.35	0.13

had only been previously recorded in the Azores archipelago on São Miguel Island (according to Borges *et al.*, 2010). This fish is native to North America and was introduced as an exotic predator in freshwater habitats on all the continents around the world, excluding Antarctica (Pyke, 2005). This small exotic fish is a voracious predator of not only small prey such as invertebrates (Walters & Legner, 1980) but also large prey such as tadpoles (e.g., Morgan & Buttemer, 1996) and other small fishes (e.g., Ivantsoff, 1999). A multiple regression analysis revealed that the electrical conductivity and abundance of *G. holbrooki* exerted significant positive and negative effects, respectively, on the abundance of *H. guernei*, accounting for 19.1% and 14.5% of the variability, respectively (Table 2). Although we detected low values of conductivity in comparison with other Mediterranean temporary ponds (e.g., Florencio *et al.*, 2013), the significant effects of these predictors on the abundance of *H. guernei* suggest a narrow species tolerance to environmental changes. Species of the genera *Hydroporus* typically occur in restrictive environments, such as acidic waters with low nutrient concentrations and *Sphagnum*-dominated vegetation (Verberk *et al.*, 2001). However, we also demonstrate that the exotic fish *G. holbrooki* can influence the occurrence and abundance of *H. guernei* in certain ponds (NEG pond, Table 1). *G.*

holbrooki generally feeds on aquatic beetle larvae and adults (Sokolov & Chvaliova, 1936; Ozturk & Ikiz, 2003; Erguden, 2013; Kalogianni *et al.*, 2014), particularly those in the family Dytiscidae (Gkenas *et al.*, 2012; Kalogianni *et al.*, 2014). Thus, we suggest that the voracious behaviour of this exotic predator fish can potentially threaten pond endemicity on these oceanic islands. However, this affirmation must be confirmed by periodic monitoring in these ponds in the future. Although we detected the exotic fish in a permanent pond, the biodiversity of temporary ponds is not exempt from risks associated with mosquito-fish invasion, as the species may eventually colonise these ponds during heavy rainfalls, when waterways may form that connect temporary and permanent ponds (e.g., Fahd *et al.*, 2007).

Implications for conservation

In our study, we demonstrate the roles of temporary ponds in the context of pond endemicity on an oceanic island. These temporary ponds are free of fish, while both permanent ponds show evidence of fish. *G. holbrooki* was only recorded in the NEG pond. Although the study included a small number of ponds, we demonstrate that permanent ponds are likely more vulnerable to predator fish invasions. Such invasions require urgent monitoring actions, including control strategies for mosquitofish. These ponds require official recognition by conservation programmes, such as the Ramsar Convention. Studies encompassing freshwaters in the Macaronesian region are scarce; thus, further research should focus on temporal and spatial dynamics in freshwater systems to establish proper guidelines for conservation (Hughes & Malmqvist, 2005).

ACKNOWLEDGMENTS

Laura Serrano performed the total phosphorous concentration analyses in the water column. Rui Elias and Rosalina Gabriel provided the technical equipment for pH, dissolved oxygen concentrations and conductivity measurements in the water column. Paulo AV Borges provided

material and support during the pond analyses on Terceira Island. We are grateful to Isabel Muñoz for comments that improved a previous version of the manuscript. We also appreciate the contributions of Marco Rosa and Adolfo Cordero for their support and comments as well as Luis An-sias, Sergio Fernández, Rui Carvalho, Herlander Lima and Will Sharkey for their assistance with fieldwork. We also thank Kieran Daniel Withey for reviewing the English. We thank the Direcção Regional da Ciência, Tecnologia e Comunicações (DRCT) for supporting the fellowship of M.F. (M3.1.7/F/002/2011) and the Fundación Barrié for the financial and academic support of L.L.L. M.F. is currently supported by the Conselho Nacional de Desenvolvimento Científico e Tecnológico-CNPq (401045/2014-5), Programa Ciência sem Fronteiras.

REFERENCES

- ALLUAUD, C. 1891. Coléoptères recueillis aux Açores par M. J. de Guerne pendant les campagnes du yacht l'Hirondelle (1887-1888). *Memoires de la Societe Zoologique de France*, 4: 197–207.
- ARRIBAS, R., C. DIAZ-PANIAGUA & I. GOMEZ-MESTRE. 2014. Ecological consequences of amphibian larvae and their native and alien predators on the community structure of temporary ponds. *Freshwater Biology*, 59(9): 1996–2008.
- BERNARD, A., C. D. ESCOBAR, A. MAZOT & R. E. GUTIÉRREZ. 2004. The acid volcanic lake of Santa Ana volcano, El Salvador. In *Natural hazards in El Salvador: Boulder, Colorado. Geological Society of America Special Paper*, 375: 121–133.
- BORGES, P. A. V., A. COSTA, R. CUNHA, R. GABRIEL, V. GONÇALVES, A. F. MARTINS, I. MELO, M. PARENTE, P. RAPOSEIRO, P. RODRÍGUEZ, R. S. SANTOS, L. SILVA, P. VIEIRA & V. VIEIRA. 2010. *A list of the terrestrial and marine biota from the Azores*. Princípiã, Cascais.
- CASSEY, P., T. M. BLACKBURN, R. P. DUNCAN & S. L. CHOWN. 2005. Concerning invasive species: reply to Brown and Sax. *Austral Ecology*, 30: 475–480.
- CLYMO, R. S. 1964. The origin of acidity in sphagnum bogs. *Briologist*, 67: 427–431.
- DELMELLE, P. & A. BERNARD. 1994. Geochemistry, mineralogy, and chemical modeling of the acid crater lake of Kawah Ijen Volcano, Indonesia. *Geochimica et Cosmochimica Acta*, 58(11): 2445–2460.
- ERGUDEN, S. A. 2013. Age, growth, sex ratio and diet of eastern mosquitofish *Gambusia holbrooki* Girard, 1859 in Seyhan Dam Lake (Adana/Turkey). *Iranian Journal of Fisheries Sciences*, 12(1): 204–218.
- FAHD, K., M. FLORENCIO, C. KELLER & L. SERRANO. 2007. The effect of the sampling scale on zooplankton community assessment in temporary ponds and its implication for the conservation of temporary ponds in south-west Spain. *Aquatic Conservation: Marine and Freshwater Ecosystem* 17(2): 175–193.
- FLORENCIO, M., L. SERRANO, C. GÓMEZ-RODRÍGUEZ, A. MILLÁN & C. DÍAZ-PANIAGUA. 2009. Inter and intra-annual variations of macroinvertebrate assemblages are related to the hydroperiod in Mediterranean temporary ponds. *Hydrobiologia*, 634: 167–183.
- FLORENCIO, M., C. DÍAZ-PANIAGUA, C. GÓMEZ-RODRÍGUEZ & L. SERRANO. 2013. Biodiversity patterns in a macroinvertebrate community of a temporary pond network. *Insect Conservation and Diversity*, 4(7): 4–21.
- GKENAS, C., A. OIKONOMOU, A. ECONOMOU, F. KIOSSE & I. LEONARDOS. 2012. Life history pattern and feeding habits of the invasive mosquitofish, *Gambusia holbrooki*, in Lake Pamvotis (NW Greece). *Journal of Biological Research-Thessaloniki*, 17: 121–136.
- GOLTERMAN, H. L. 2004. *The chemistry of phosphate and nitrogen compounds in sediments*. Kluwer Academic Publisher, Dordrecht, The Netherlands.
- HEYER, W. R., M. A. DONNELLY, R. W. MCDIARMID, L.-A. C. HAYEK & M. S. FOSTER. 1994. *Measuring and monitoring biological diversity. Standard methods for amphibians*. Smithsonian Institution Press, Washington.
- HERMOSO, V., M. CLAVERO & M. J. KENNARD. 2012. Determinants of fine-scale homogenization and differentiation of native freshwater fish faunas in a Mediterranean Basin: implications for conservation. *Diversity and Distributions*, 18: 236–247.
- HUGHES, S. J. & B. MALMQVIST. 2005. Atlantic Island freshwater ecosystems: challenges and considerations following the EU Water Framework Directive. *Hydrobiologia*, 544(1): 289–297.

- IVANTSOFF, W. 1999. Detection of predation on Australian native fishes by *Gambusia holbrooki*. *Marine and Freshwater Research*, 50(5): 467–468.
- KALOGIANNI, E., S. GIAKOUMI, A. ANDRIOPOULOU & Y. CHATZINIKOLAOU. 2014. Prey utilisation and trophic overlap between the non native mosquitofish and a native fish in two Mediterranean rivers. *Mediterranean Marine Science*, 15(2): 287–301.
- LEGENDRE, P. & L. LEGENDRE (1998) *Numerical ecology: developments in environmental modelling* 20, 2nd ed. Elsevier, New York.
- LEPS, J. & P. SMILAUER (2003) *Multivariate Analysis of Ecological Data Using CANOCO*. Cambridge University Press, New York City, New York.
- MENDES, C. & E. DIAS. 2001. *Ecologia e Vegetação das Turfeiras de Sphagnum spp. da ilha Terceira (Açores)*. Comunidade Europeia, Direcção Regional dos Recursos Florestais, Universidade dos Açores, Angra do Heroísmo.
- MORGAN, L. A. & W. A. BUTTEMER. 1996. Predation by the non-native fish *Gambusia holbrooki* on small *Litoria aurea* and *L. dentata* tadpoles. *Australian Zoologist*, 30(2): 143–149.
- OZTURK, S. & R. IKIZ. 2003. Some biological properties in the Akgöl (Fethiye-Muğla) population of the mosquitofish *Gambusia affinis* (Baird and Girard, 1853). *Turkish Journal of Veterinary and Animal Science*, 27: 911–915.
- PYKE, G. H. 2005. A review of the biology of *Gambusia affinis* and *G. holbrooki*. *Reviews in Fish Biology and Fisheries*, 15(4): 339–365.
- RAPOSEIRO, P. M., S. J. HUGHES, A. C. COSTA & A. MAFALDA CRUZ. 2012. Azorean freshwater invertebrates: status, threats and biogeographic notes. *Limnetica*, 31(1): 13–22.
- SOKOLOV, N. P. & M. A. CHVALIOVA. 1936. Nutrition of *Gambusia affinis* on the rice fields of Turkestan. *The Journal of Animal Ecology*, 5(2): 390–395.
- TIMPERLEY, M. H. & R. J. VIGOR & BROWN. 1986. Water chemistry of lakes in the Taupo Volcanic Zone, New Zealand. *New Zealand Journal of Marine and Freshwater Research*, 20(2): 173–183.
- VERBERK, W. C. E. P., G. J. A. VAN DUINEN, T. M. PEETERS & H. ESSELINK. 2001. Importance of variation in water-types for water beetle fauna (Coleoptera) in Korenburgerveen, a bog remnant in the Netherlands. *Proceedings of the Section Experimental and Applied Entomology-Netherlands Entomological Society*, 12: 121–128.
- VITOUSEK, P. M., C. M. D'ANTONIO, L. L. LOOPE & R. WESTBROOKS. 1996. Biological invasions as global environmental change. *American Scientist*, 84: 468–478.
- VITOUSEK, P. M., H. A. MOONEY, J. LUBCHENCO & J. M. MELILLO. 1997. Human domination of Earth's ecosystems. *Science*, 277: 494–499.
- WALTERS, L. L. & E. F. LEGNER. 1980. Impact of the desert pupfish *Cyprinodon macularius*, and *Gambusia affinis affinis* on fauna in pond ecosystems. *Hilgardia*, 48(3): 1–18.
- WELLBORN, G. A., D. K. SKELLY & E. E. WERNER. 1996. Mechanisms creating community structure across a freshwater habitat gradient. *Annual review of ecology and systematics*, 27(1): 337–363.
- WHITTAKER, R. J. & J. M. FERNÁNDEZ-PALACIOS. 2007. *Island biogeography: ecology, evolution, and conservation*, 2nd ed. Oxford University Press, Oxford.
- WILLIAMS, D. D. 2006. *The Biology of Temporary Waters*. Oxford: Oxford University Press.